

# Greenhouse Gas Emissions from Rice Production in Arkansas

**Kristofor R. Brye**  
Professor – Applied Soil  
Physics and Pedology

**Diego Della Lunga**  
Graduate Research  
Assistant

**Trent L. Roberts**  
Professor –  
Soils Specialist

## Importance of Greenhouse Gases

Climate change and climate change mitigation are important environmental and agronomic issues. One factor responsible for climate change is increasing greenhouse gas (GHG) concentrations in the atmosphere.

The major greenhouse gases are carbon dioxide (CO<sub>2</sub>), methane (CH<sub>4</sub>), nitrous oxide (N<sub>2</sub>O), ozone (O<sub>3</sub>) and water vapor (H<sub>2</sub>O). In an agricultural setting, CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O are the main GHGs of concern. Microbial and plant root respiration are the primary sources of carbon dioxide emissions associated with agricultural crop production.

Carbon dioxide is the by-product of the aerobic breakdown of organic matter by microorganisms in the soil and is released from plant roots into the rhizosphere. Methane is produced via methanogenesis from the microbial decomposition of organic matter in the soil, but under anaerobic (i.e., lacking oxygen) soil conditions. Nitrous oxide is produced from the incomplete conversion of soil nitrate (NO<sub>3</sub><sup>-</sup>) to dinitrogen gas (N<sub>2</sub>) in the process of denitrification, which can occur during wet and dry cycles, commonly under oxygen-limited soil conditions.

Because concentrations of CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O are not large in the atmosphere, they are referred to as trace gases (Smartt et al., 2016a).

They can, however, trap outgoing long-wave radiation (i.e., heat energy) emitted from the Earth's surface. Preventing the escape of long-wave radiation increases the temperature of the Earth's atmosphere, just as a greenhouse traps heat inside a structure for plant growth.

Even small increases in the atmospheric air temperature can disrupt many global processes. The melting of glaciers and ice caps has resulted in sea-level rise, for example, and shifting weather patterns have caused increased the frequency drought, flooding and other extreme conditions.

Collectively, the sum of CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O concentrations in the atmosphere are used to calculate the global warming potential (GWP), as reported in CO<sub>2</sub> equivalents. Individually, CH<sub>4</sub> and N<sub>2</sub>O are both more potent GHGs than CO<sub>2</sub>, with GWPs ~ 25 to 30 and ~ 265 to 300 times greater, respectively, than that of CO<sub>2</sub> alone (Smartt et al., 2016a).

Other typical quantifications of GHGs include GHG flux (mass per unit area per unit time) and GHG emissions (mass per unit area summed over a specific period). Understanding how various production systems and associated management practices influence GHG emissions is critical for sustainable agriculture. Studies on GHG fluxes and emissions in agricultural settings can also provide information to help guide management decisions

*Arkansas Is  
Our Campus*

Visit our website at:  
<https://www.uaex.uada.edu>

to maintain and improve soil health, as C and N are closely connected to soil health goals, but C and N emissions represent potentially major loss pathways from ecosystems.

## Greenhouse Gas Emissions in Rice Production

Rice production in Arkansas occupies a substantial agricultural land area (typically between 1 million and 1.5 million acres each year) that is critical for the social and economic systems in Arkansas, collectively supplying jobs, income, and food for tens of thousands of residents. However, the U.S. Environmental Protection Agency has identified rice production as a significant contributor of GHG emissions, namely CH<sub>4</sub>, from the traditional direct-seeded, delayed, continuous flood-irrigation water management scheme (USEPA, 2011) that has been the dominant water management scheme for rice production in Arkansas for decades.

Flood-irrigation for rice production produces anaerobic soil conditions within just a few weeks of flood application (Brye et al., 2013), making traditional flood-irrigated rice production uniquely susceptible to CH<sub>4</sub> production via methanogenesis and the release of CH<sub>4</sub> to the atmosphere. Under flood-irrigated rice, the main conduit from CH<sub>4</sub> emissions to the atmosphere is through specialized cells in the rice plants themselves. Secondary pathways for CH<sub>4</sub> from the soil to the atmosphere under flood-irrigated rice include diffusion and ebullition (bubbles) through the floodwater.

Though the development of saturated, anaerobic soil conditions under flood-irrigated rice increases CH<sub>4</sub> production and release, application of a flood for rice production greatly reduces CO<sub>2</sub> and N<sub>2</sub>O production and emissions. However, due to the relatively shallow flood-water layer commonly used in rice production systems in the United States (Henry et al., 2021a; Liu and Wu, 2004; Vories et al., 2002), CO<sub>2</sub> and N<sub>2</sub>O emissions are never completely eliminated.

Several alternative water management schemes use significantly less water than traditional, continuous flood-irrigated rice production. Alternative practices, such as alternate wetting and drying (AWD), intermittent flooding and the more recent wide-scale implementation of furrow-irrigated rice (row rice), have been shown to reduce overall water use (He, 2010; Henry et al., 2021a). These alternatives also provide some economic savings from reduced labor and energy costs compared to traditional flood-irrigated rice production, which requires the establishment and maintenance of levees to contain flood water (Henry et al., 2021a,b).

These alternate schemes have their own environmental consequences, however. The greater frequency of wet-and-dry soil cycles will reduce CH<sub>4</sub> production, but will generally stimulate greater CO<sub>2</sub> and N<sub>2</sub>O production and emissions. Several important soil/plant characteristics and agricultural management practices in Arkansas are known to affect GHG emissions, particularly CH<sub>4</sub>.

Soil texture (i.e., silt-loam versus clay soil) affects CH<sub>4</sub> emissions, where the fine-textured clay soils generally have lower CH<sub>4</sub> emissions than coarser-textured silt-loam soils, which reduce more quickly and have a greater proportion of larger pores, allowing gas to escape the soil more quickly (Brye et al., 2013).

Cultivar selection (i.e., pure-line versus hybrid) affects CH<sub>4</sub> emissions, as hybrid cultivars typically have more vigorous growth and greater biomass production than pure-line cultivars, which allowed for a more oxygenated rhizosphere to convert CH<sub>4</sub> back to CO<sub>2</sub> in the soil before escaping to the atmosphere (Rogers et al., 2014).

Fertilizer-nutrient source (i.e., inorganic versus organic), particularly nitrogen source, affects CH<sub>4</sub>. Organic fertilizers, such as poultry litter, provide a readily available source of carbon in the soil that can be converted to CH<sub>4</sub> upon the decomposition of the organic material — this doesn't happen with inorganic fertilizer-nutrient sources (Rogers et al., 2017).

Poultry litter as a fertilizer-nutrient source or general organic soil amendment will also increase CO<sub>2</sub> emissions (Brye et al., 2006). Since rice is generally considered a high-residue-producing crop, the amount of carbon-containing plant material that is recycled back to the soil each year can be quite high. Therefore, the previous crop and/or crop rotation (e.g., rice-soybean versus rice-rice rotation) can affect CH<sub>4</sub> emissions, where emissions are generally larger when rice is grown in consecutive years, compared to fields in which soybean was the previous crop. Soybean is generally a low-residue-producing crop that returns much less readily decomposable C to the soil than does a rice crop (Rogers et al., 2014).

Additionally, when rice follows rice, the soil tends to stay more reduced and is quicker to become further reduced in the following rice crop, thus allowing CH<sub>4</sub> emissions to occur sooner.

Conventional tillage has been shown to affect N<sub>2</sub>O emissions from furrow-irrigated rice (Slayden et al., 2022a), where N<sub>2</sub>O emissions were greater from conventional tillage than from no-tillage, while tillage has been shown to have little effect on CH<sub>4</sub>

emissions from both flood-irrigated (Rector et al., 2018a) or furrow-irrigated (Della Lunga et al., 2023) rice production. However, Della Lunga et al. (2023) demonstrated that GWP tended to be lower from no-tillage compared to conventional tillage in furrow-irrigated rice.

## Summary of Research

Since 2013, numerous field and greenhouse studies have evaluated various soil, plant and agricultural effects on GHG emissions from rice production in Arkansas. Some studies have measured CO<sub>2</sub>, CH<sub>4</sub>, or N<sub>2</sub>O only, while others have measured multiple trace gases. Except for the CO<sub>2</sub>-only studies, GHG research has been conducted using the vented, non-flow-through, non-steady-state closed-chamber approach (Parkin and Venterea, 2010), with chamber diameters of 15 or 30 cm (Figures 1 and 2).

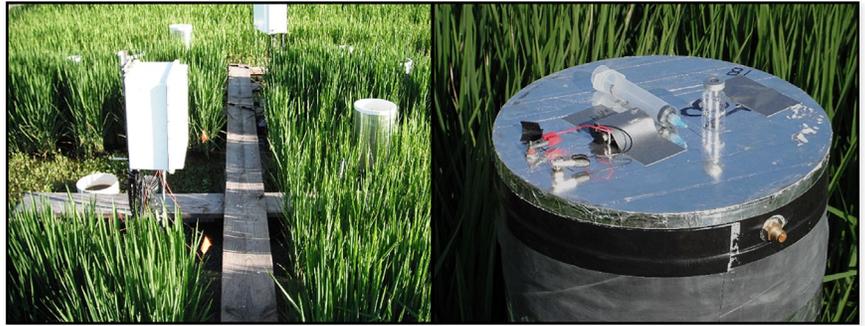
### CO<sub>2</sub> Only Studies

Brye et al. (2006) evaluated soil CO<sub>2</sub> fluxes from pelletized and fresh poultry litter before flooding rice on silt-loam soil. Results showed that soil CO<sub>2</sub> fluxes were generally unaffected by poultry litter form, but generally increased as the litter application rate increased (Brye et al., 2006). Motschenbacher et al. (2015) evaluated tillage and crop rotation effects on soil CO<sub>2</sub> fluxes during the non-flooded period of flood-irrigated rice on silt-loam soil. After more than 10 years of consistent management, few differences in early and late-season soil CO<sub>2</sub> fluxes occurred between tillage treatments and among various crop rotations including rice (Motschenbacher et al., 2015).

### CH<sub>4</sub> Only Studies

Rogers et al. (2013) evaluated CH<sub>4</sub> fluxes over time with and without plants in the measurement chamber (Figure 1), with chambers both in and between rows and with optimal and no-N fertilization in drill-seeded, delayed-flood rice on silt-loam soil. Results showed that CH<sub>4</sub> emissions were greater with than without plants, but unaffected by N fertilization (Rogers et al., 2013). Brye et al. (2013) summarized the effects of soil texture on CH<sub>4</sub> fluxes from optimally N-fertilized rice grown in drill-seeded, delayed-flood rice on a silt-loam and clay soil, where season-long CH<sub>4</sub> emissions from silt-loam soil were greater than from clay soil.

Rogers et al. (2014) evaluated the effects of rotation (soybean-rice and rice-rice) and cultivar (pure-line and hybrid) on season-long methane emissions in drill-seeded, delayed-flood rice on silt-loam soil.



**Figure 1.** Greenhouse gas emissions field research set-up for flood-irrigated rice, out-fitted with boardwalks between plots to minimize soil disturbance, 30-cm diameter chambers, sensors, and a data logger to continuously measure and record soil temperatures and soil oxidation-reduction (redox) potentials (left panel). Chamber cap used to collect gas samples from the chamber headspace, which was mixed with a small, 9-V-battery-operated fan, using a syringe poked through a septum (right panel).

Results showed that season-long CH<sub>4</sub> emissions were greater from a rice-rice than a rice-soybean rotation and greater from a pure-line than from a hybrid cultivar (Rogers et al., 2014). Smartt et al. (2016b) evaluated the effects of N-fertilized, non-N-fertilized and bare-soil treatments on CH<sub>4</sub> emissions from drill-seeded, delayed-flood rice on clay soil. Results showed that season-long CH<sub>4</sub> emissions were greater when rice was fertilized with N. Brye et al. (2016) summarized the effects of soil texture, rice cultivar and crop rotation on CH<sub>4</sub> emissions, where results showed that season-long CH<sub>4</sub> emissions were greater from silt-loam than clay soils, greater from a pure-line than a hybrid cultivar and greater from a rice-rice than a rice-soybean rotation.

Rogers et al. (2017) evaluated the effects of fertilizer-N source (ammonium sulfate, urea only and pelletized poultry litter plus urea), previous crop (soybean and rice) and soil texture (silt loam and clay) on CH<sub>4</sub> emissions from drill-seeded, delayed-flood rice. Results showed that season-long CH<sub>4</sub> emissions were greater from pelletized poultry litter than from ammonium sulfate or urea, greater from a silt-loam than clay soil and greater when rice was the previous crop compared to soybean on the clay soil, but were unaffected by previous crop on the silt-loam soil (Rogers et al., 2017).

Brye et al. (2017a) evaluated the diurnal variation in CH<sub>4</sub> fluxes and emissions during vegetative and reproduction growth stages in a silt-loam and clay soil, where results showed that CH<sub>4</sub> fluxes differed over time throughout the day and were greater from a pure-line than a hybrid cultivar on silt-loam soil, but not on clay soil. Brye et al. (2017b) evaluated the effects of rice cultivar selection (one hybrid and two pure-line cultivars) and crop rotation (rice-rice and

soybean-rice) on CH<sub>4</sub> fluxes and emissions across multiple years from drill-seeded, delayed-flood rice production. Results showed that season-long CH<sub>4</sub> emissions were greater from the two pure-line cultivars than from the hybrid and greater following rice than soybean (Brye et al., 2017b).



Smartt et al. (2018) evaluated the effects of three hybrid (CLXL729, CLXL745, and XL753) and a pure-line (Roy J) rice cultivar on CH<sub>4</sub> fluxes and seasonal emissions from silt-loam soil. Results showed that season-long CH<sub>4</sub> emissions were greater from the pure-line than from any of the three hybrid cultivars (Smartt et al., 2018). Humphreys et al. (2018a) evaluated the effects of water management scheme (full-season flood and mid-season drain) on season-long CH<sub>4</sub> emissions in the direct-seeded, delayed-flood rice production system on silt-loam soil. Results showed that season-long CH<sub>4</sub> emissions were greater from the full-season-flood than from the mid-season-drain water management scheme.

Humphreys et al. (2018b) evaluated the effects of tillage (conventional tillage and no-tillage) and coated and uncoated urea on CH<sub>4</sub> fluxes in the direct-seeded, delayed-flood rice production system on silt-loam soil, where results showed that season-long CH<sub>4</sub> emissions were unaffected by tillage or urea type. Humphreys et al. (2018c) evaluated the effects of soil organic matter (SOM) concentration on CH<sub>4</sub> emissions from direct-seeded, delayed-flood rice in silt-loam soils, where results showed that season-long CH<sub>4</sub> emissions increased as initial SOM concentration increased.

### **N<sub>2</sub>O Only Studies**

Rector et al. (2018a) evaluated the effects of tillage practice (conventional tillage and no-tillage) and urea-fertilizer type (coated and uncoated) on N<sub>2</sub>O fluxes and season-long emissions from direct-seeded, delayed-flood rice on silt-loam soil, where results showed that neither tillage nor urea-fertilizer type affected season-long N<sub>2</sub>O emissions. Rector et al. (2018b) evaluated the effects of water management scheme (full-season flood and intermittent flood) and rice cultivar (pure-line and hybrid) on N<sub>2</sub>O fluxes and season-long emissions from direct-seeded, delayed-flood rice on silt-loam soil, where results showed that neither water management scheme nor rice cultivar affected season-long N<sub>2</sub>O emissions.

Slayden et al. (2022a) evaluated the effects of site position (up-, middle- and down-slope) and tillage (conventional tillage and no-tillage) on N<sub>2</sub>O emissions in furrow-irrigated rice on silt-loam soil (Figure 2). Results showed that season-long N<sub>2</sub>O emissions were



**Figure 2. Greenhouse gas emissions field research set-up for furrow-irrigated rice, out-fitted with 30-cm diameter chambers, sensors, and a data logger to continuously measure and record soil temperatures and soil oxidation-reduction (redox) potentials at one site position in a production-scale field (left panel). Greenhouse gas chambers with caps and extenders to accommodate growing rice later in the growing season (right panel).**

greater from the down-slope than the up-slope or mid-slope positions and greater from conventional tillage than no-tillage (Slayden et al., 2022a).

Slayden et al. (2022b) evaluated the effects of fertilizer-N rates and application timings on N<sub>2</sub>O fluxes and season-long emissions in simulated furrow-irrigated rice grown on silt-loam soil in a greenhouse. Results showed that season-long N<sub>2</sub>O emissions were unaffected by fertilizer-N rate or application timing, but that numeric reductions in season-long N<sub>2</sub>O emissions could be achieved with split N applications. In general, season-long N<sub>2</sub>O emissions are typically several orders of magnitude lower than season-long CH<sub>4</sub> emissions across rice production practices and soil characteristics.

### **CO<sub>2</sub>, CH<sub>4</sub>, and/or N<sub>2</sub>O Combined Studies**

Della Lunga et al. (2020) evaluated the effects of soil moisture status (flooded; saturated, but not flooded; and moist soil, slightly below saturation) on season-long CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from rice grown on silt-loam soil in the greenhouse. Results showed that season-long CO<sub>2</sub> and CH<sub>4</sub> emissions were greater from the flooded-soil condition than the other water management regimes, while season-long N<sub>2</sub>O emissions were unaffected by water management regime (Della Lunga et al., 2020).

Della Lunga et al. (2021a) evaluated the relationships among soil moisture, oxidation-reduction (redox) potential and soil temperature and CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O fluxes and emissions across the elevation gradient in a furrow-irrigated rice field on silt-loam soil. Results showed that CO<sub>2</sub> fluxes were directly correlated with soil moisture and temperature, but inversely correlated with soil redox potential; CH<sub>4</sub>

fluxes were directly correlated with soil moisture and inversely correlated with soil redox potential; and N<sub>2</sub>O fluxes were directly correlated with soil redox potential (Della Lunga et al., 2021a).

Della Lunga et al. (2021b) evaluated the effects of site position (up-, middle- and down-slope) and tillage practice (conventional tillage and no-tillage) on diurnal variations of CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O fluxes during vegetative rice growth in a furrow-irrigated rice field on silt-loam soil. Results showed that daily CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O flux maxima occurred in the late afternoon, while flux minima occurred in the early morning (Della Lunga et al., 2021b).

Della Lunga et al. (2023) evaluated the effects of site position (up-, middle- and down-slope) and tillage

(conventional tillage and no-tillage) on CO<sub>2</sub> and CH<sub>4</sub> emissions and GWP in furrow-irrigated rice on silt-loam soil. Results showed that season-long GWP was generally greater from conventional tillage than from no-tillage and was greater at the down-slope than at the up-slope or mid-slope positions (Della Lunga et al., 2023).

## Recommendations to Minimize Greenhouse Gas Emissions from Rice Production

Agricultural decisions involving soil/plant characteristics and management practices can often be adjusted to reduce a single GHG (Table 1). However, it is more challenging to prescribe a course of action that will simultaneously reduce multiple GHGs.

**Table 1. Summary of major known soil and agronomic factors affecting greenhouse gas (GHG) emissions from rice production, the specific soil/plant characteristic or management practice to choose to minimize methane (CH<sub>4</sub>) or nitrous oxide (N<sub>2</sub>O) emissions, and the relevant reference(s) to support the recommendation.**

| Major Factor                             | Specific Soil/Plant Characteristic or Practice | Characteristic/Practice to Minimize Specific GHG Emissions |                  | Relevant References  |
|--|--|--|------------------|--|
|  |  | CH <sub>4</sub>  | N <sub>2</sub> O |  |
| Soil texture                             | Silt loam                                      |  | -†               | Brye et al. (2013), Brye et al. (2016)<br>Rogers et al. (2017)   |
|  | Clay   | X  | -                |  |
| Cultivar selection                       | Pure-line                                      |  | -                | Rogers et al. (2014), Brye et al. (2016), Brye et al. (2017), Smartt et al. (2018), Rector et al (2018b)                         |
|  | Hybrid   | X  | -                |  |
| Nutrient source                          | Inorganic                                      | X  | -                | Rogers et al. (2017)   |
|  | Organic  |  | -                |  |
| Nutrient application timing              | Optimum N + one split                          | -  |                  | Slayden et al. (2022b)   |
|  | Half optimum N + two splits                    | -  | X                |  |
|  | Optimum N + two splits                         | -  | X                |  |
| Water management scheme                  | Flood irrigation                               |  | X                | Humphreys et al. (2018a), Rector et al. (2018b)  |
|  | Non-flood irrigation                           | X  |                  |  |
| Previous crop                            | Soybean  | X  | -                | Rogers et al. (2014), Brye et al. (2016), Brye et al. (2017)   |
|  | Rice   |  | -                |  |
| Tillage                                  | Conventional tillage                           | -  |                  | Rector et al. (2018a), Humphreys et al. (2018b), Slayden et al. (2022a)<br>Della Lunga et al. (2021b), Della Lunga et al. (2023) |
|  | No-tillage                                     | -  | X                |  |
| Soil moisture condition                  | Flooded  |  | -                | Della Lunga et al. (2020)  |
|  | Saturated, but not flooded                     | X  | -                |  |
|  | Moist, near saturation, but not saturated      | X  | -                |  |
| Soil moisture and temperature variations | Small variation                                | -  | X                | Della Lunga et al. (2020)  |
|  | Large variation                                | -  |                  |  |

† Dashes indicate either little measured effect from the specific characteristic or practice, or the research has not been conducted yet to confirm any effect.

Based on the results of field and greenhouse research studies conducted in Arkansas, here are some recommendations for reducing individual GHGs and multiple GHGs simultaneously:

## CO<sub>2</sub>

- Decreased tillage intensity and/or frequency will contribute to reduced CO<sub>2</sub> emissions.
- Cover crop and/or crop residue presence will reduce soil temperature and moisture fluctuations, reducing the rate of soil respiration and lowering CO<sub>2</sub> emissions.
- Maintaining uniform soil moisture conditions or avoiding large and frequent soil moisture fluctuations in furrow-irrigated rice production will contribute to reduced CO<sub>2</sub> emissions.

## CH<sub>4</sub>

- Flood-irrigated rice production on clay soils will result in reduced CH<sub>4</sub> emissions compared to silt-loam or coarser-textured soils.
- Flood-irrigated rice production with a hybrid rice cultivar will contribute to reduced CH<sub>4</sub> emissions.
- Using inorganic fertilizer-nutrient sources in flood-irrigated rice production will contribute to reduced CH<sub>4</sub> emissions.
- Flood-irrigated rice production in rotation with soybean will contribute to reduced CH<sub>4</sub> emissions.
- Non-flood-irrigated rice production techniques such as AWD, intermittent flooding and furrow-irrigation, will contribute to reduced CH<sub>4</sub> emissions.

## N<sub>2</sub>O

- Flood-irrigated rice production will contribute to reduced N<sub>2</sub>O emissions.
- Furrow-irrigated rice production with no-tillage will contribute to reduced N<sub>2</sub>O emissions.
- Maintaining uniform soil moisture conditions or avoiding large and frequent soil moisture fluctuations in furrow-irrigated rice production will contribute to reduced N<sub>2</sub>O emissions.
- Split applications of inorganic N fertilizer will contribute to reduced N<sub>2</sub>O emissions.

## CO<sub>2</sub>, CH<sub>4</sub> and/or N<sub>2</sub>O Combined

- Maintaining uniform soil moisture conditions or avoiding large and frequent soil moisture fluctuations in furrow-irrigated rice production

will contribute to reduced CO<sub>2</sub> and N<sub>2</sub>O emissions.

- Flood-irrigated rice production on clay soils contributes to reduced CH<sub>4</sub> and N<sub>2</sub>O emissions.
- Furrow-irrigated rice production with no-tillage will contribute to reduced CH<sub>4</sub> and N<sub>2</sub>O emissions.
- Furrow irrigation can reduce GWP compared to flood-irrigated conditions.

## Conclusions

Greenhouse gas emissions from Arkansas' rice-producing soils, most importantly silt-loam soils, are a concern for conservation planning and for maintaining sustainable use of the state's soil, water and air resources. As one of the major land uses in Arkansas and a significant contributor to Arkansas' agricultural economy, traditional flood-irrigated rice has been specifically identified as a major source of GHG emissions, namely CH<sub>4</sub>.

Field research in Arkansas has confirmed soil texture and water management scheme are the most important factors affecting GHG emissions, namely CH<sub>4</sub> and N<sub>2</sub>O, from rice production. Flood-irrigated rice production on silt-loam soils is much more of a concern than on clay soils. While being most widespread, flood-irrigated rice production, especially on silt-loam soils, produces greater CH<sub>4</sub> emissions than non-flood-irrigated rice production or clay soils.

Despite greater CH<sub>4</sub> emissions from flood-irrigated rice on silt-loam soil than from non-flood-irrigated rice production or from clay soils, season-long GHG emissions, namely CH<sub>4</sub>, from direct field measurements from rice production in Arkansas are generally lower than what EPA has suggested for season-long emissions from rice production in general.

Furthermore, direct field measurements of season-long CH<sub>4</sub> emissions from rice production in Arkansas are generally lower than those reported for many other rice production systems around the world, particularly deep-flooded paddy rice. Reducing GHG emissions would be a significant conservation measure that would benefit Arkansas' agricultural producers and landowners in general, not just rice producers.

Although GHGs are being emitted, it appears that Arkansas rice production generates less GHG than once thought as compared to other rice-producing

states and countries. Nevertheless, continued research regarding rice production and GHG emissions is necessary to conserve Arkansas' valuable soil, water and air resources.

## References:

- Brye, K.R., A.D. Smartt, and R.J. Norman. 2017a. Diurnal methane fluxes as affected by cultivar from direct-seeded, delayed-flood rice production. *J. Environ. Prot.* 8:957-973.
- Brye, K.R., B. Golden, and N.A. Slaton. 2006. Poultry litter decomposition dynamics as affected by litter form and rate prior to flooding for rice production. *Soil Sci. Soc. Am. J.* 70:1155-1167.
- Brye, K.R., C.W. Rogers, A.D. Smartt, and R.J. Norman. 2013. Soil texture effects on methane emissions from direct-seeded, delayed-flood rice production in Arkansas. *Soil Sci.* 178:519-529.
- Brye, K.R., C.W. Rogers, A.D. Smartt, R.J. Norman, J.T. Hardke, and E.E. Gbur. 2017b. Methane emissions as affected by crop rotation and rice cultivar in the Lower Mississippi River Valley, USA. *Geoderma Reg.* 11:8-17.
- Brye, K.R., L.L. Nalley, J.B. Tack, B.L. Dixon, A.P. Barkley, C.W. Rogers, A.D. Smartt, R.J. Norman, and K. Jagadish. 2016. Factors affecting methane emissions from rice production in the Lower Mississippi River Valley, USA. *Geoderma Reg.* 7:223-229.
- Della Lunga, D., K.R. Brye, T.L. Roberts, and S.G. Lebeau. 2020. Water management effects on trace gas emissions under greenhouse conditions from direct-seeded hybrid rice in silt-loam soil. *J. Rice Res. Dev.* 3:95-102.
- Della Lunga, D., K.R. Brye, J.M. Slayden, C.G. Henry, and L.S. Wood. 2021a. Relationships among soil factors and greenhouse gas emissions from furrow-irrigated rice in the mid-southern, USA. *Geoderma Reg.* 24:e00365.
- Della Lunga, D., K.R. Brye, J.M. Slayden, C.G. Henry, and S. Park. 2021b. Diurnal variation of trace gas fluxes during the vegetative lag phase in furrow-irrigated rice production. *J. Rice Res. Dev.* 4:338-350.
- Della Lunga, D., K.R. Brye, J.M. Slayden, and C.G. Henry. 2023. Evaluation of site position and tillage effects on global warming potential from furrow-irrigated rice in the mid-southern USA. *Geoderma Reg.* 32:e00625.
- He, C. 2010. Effects of furrow irrigation on the growth, production, and water use efficiency of direct sowing rice. *Sci. World J.* 10:1483-1497.
- Henry, C.G., J.P. Pimentel, P.N. Gahr, and T. Clark. 2021a. Evaluating irrigation timing, depletion, water-use and efficiencies in furrow irrigated-rice. pp. 255-258. In: J. Hardke, X. Sha, and N. Bateman, Editors. *B.R. Wells Arkansas Rice Research Studies 2020. Research Series 676*, University of Arkansas System, Division of Agriculture, Arkansas Agricultural Experiment Station, Fayetteville.
- Henry, C.G., G.D. Simpson, R. Mane, J.P. Pimentel, and T. Clark. 2021b. Water use and yield differences in farmer-managed furrow irrigated and multiple inlet rice irrigation levee flooded fields. pp. 247-249. In: J. Hardke, X. Sha, and N. Bateman, Editors. *B.R. Wells Arkansas Rice Research Studies 2020. Research Series 676*, University of Arkansas System, Division of Agriculture, Arkansas Agricultural Experiment Station, Fayetteville.
- Humphreys, J., K.R. Brye, C. Rector, E.E. Gbur, and J.T. Hardke. 2018a. Water management and cultivar effects on methane emissions from direct-seeded, delayed-flood rice production in Arkansas. *J. Rice Res. Dev.* 1:14-24.
- Humphreys, J.J., K.R. Brye, C. Rector, E.E. Gbur, and N.A. Slaton. 2018b. Methane production as affected by tillage practice and NBPT rate from silt-loam soil in Arkansas. *J. Rice Res. Dev.* 1:49-58.
- Humphreys, J.J., K.R. Brye, C. Rector, and E.E. Gbur. 2018c. Methane emissions from rice fields across a soil organic matter gradient in Alfisols of Arkansas, USA. *Geoderma Reg.* 16:e00200.
- Liu, C., and C. Wu. 2004. Evaluation of methane emissions from Taiwanese paddies. *Sci. Tot. Environ.* 333:193-207.
- Motschenbacher, J., K. Brye, M. Anders, E. Gbur, N. Slaton, and M. Evans-White. 2015. Daily soil surface CO<sub>2</sub> flux during non-flooded periods in flood-irrigated rice rotations. *Agron. Sustain. Dev.* 35:771-782.
- Parkin, T., and R. Venterea. 2010. Chamber-based trace gas flux measurements. <http://www.ars.usda.gov/SP2UserFiles/Program/212/Chapter%203.%20GRACEnet%20Trace%20Gas%20Sampling%20Protocols.pdf> (verified 28 April, 2023).
- Rector, C., K.R. Brye, J. Humphreys, R.J. Norman, N.A. Slaton, E.E. Gbur, C. Willett, and M.A. Evans-White. 2018a. Tillage and coated-urea effects on nitrous oxide emissions from direct-seeded, delayed-flood rice production in Arkansas. *J. Rice Res. Dev.* 1:25-37.
- Rector, C., K.R. Brye, J. Humphreys, R.J. Norman, E.E. Gbur, J.T. Hardke, C. Willett, and M.A. Evans-White. 2018b. N<sub>2</sub>O emissions and global warming potential as affected by water management and rice cultivar on an Alfisol in Arkansas, USA. *Geoderma Reg.* 14:e00170.

- Rogers, C.W., A.D. Smartt, K.R. Brye, and R.J. Norman. 2017. Nitrogen source effects on methane emissions from drill-seeded, delayed-flood rice production. *Soil Sci.* 182:9-17.
- Rogers, C.W., K.R. Brye, R.J. Norman, E.E. Gbur, J.D. Mattice, T.B. Parkin, and T.L. Roberts. 2013. Methane emissions from drill-seeded, delayed-flood rice production on silt-loam soil in Arkansas. *J. Environ. Qual.* 42:1059-1069.
- Rogers, C.W., K.R. Brye, A.D. Smartt, R.J. Norman, E.E. Gbur, and M.E. Evans-White. 2014. Cultivar and previous crop effects on methane emissions from drill-seeded, delayed-flood rice production on silt-loam soil. *Soil Sci.* 179:28-36.
- Slayden, J.M., K.R. Brye, D. Della Lunga, C.G. Henry, L.S. Wood, and D.J. Lessner. 2022a. Site position and tillage treatment effects on nitrous oxide emissions from furrow-irrigated rice on a silt-loam Alfisol in the Mid-south, USA. *Geoderma Reg.* 28:e00491.
- Slayden, J.M., K.R. Brye, and D. Della Lunga. 2022b. Nitrogen fertilizer application timing effects on nitrous oxide emissions from simulated furrow-irrigated rice on silt-loam soil in the greenhouse. *J. Rice Res. Dev.* 5:366-377.
- Smartt, A.D., K.R. Brye, and R.J. Norman. 2016a. Methane emissions from rice production in the United States: A review of controlling factors and summary of research. pp. 179-207. In: *Greenhouse Gases*. B. Llamas and J. Pous, Eds., InTech, Rijeka, Croatia, ISBN 978-953-51-4323-9.
- Smartt, A.D., K.R. Brye, and R.J. Norman. 2018. Methane emissions among hybrid rice cultivars in the mid-southern United States. *Annals Adv. Agric. Sci.* 2:1-13.
- Smartt, A.D., K.R. Brye, C.W. Rogers, R.J. Norman, E.E. Gbur, J.T. Hardke, and T.L. Roberts. 2016b. Characterization of methane emissions from rice production on a clay soil in Arkansas. *Soil Sci.* 181:57-67.
- United States Environmental Protection Agency (USEPA). 2011. Inventory of U.S. greenhouse gas emissions and sinks: 1990–2009 [On-line]. Available at: <https://www.epa.gov/ghgemissions/inventory-us-greenhouse-gas-emissions-and-sinks-1990-2009> (verified 28 April, 2023).
- Vories, E., P. Counce, and T. Keisling. 2002. Comparison of flooded and furrow irrigated rice on clay. *Irrig. Sci.* 21:139-144.